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societies [4], and economic systems [5] work is often reminiscent of quantum physical phenomena. This highlights synergies that can be exploited for genuinely novel interdisciplinary research [6].

Ecologists – and researchers from other fields – have already begun to use quantum physics as an analogous model for describing patterns and processes in ecosystems including species diversity and distribution modeling [7, 8], the quantification of evolutionary processes [9], dynamic management and conservation of nature reserves [10], modeling of desertification [11], and management for sustainability [12]. Inspired by Erwin Schrödinger's famous cat that is simultaneously dead and alive [13], work has also envisioned the potential to create quantum superposition (i.e., particles occur at several places or in different states at the same time) experimentally in viruses and microorganisms such as tardigrades [14]. While the experimental induction of superposition states for living organisms is currently elusive, we will show and discuss the widespread but hitherto ignored occurrence of superposition in statistical modeling of plant and animal communities frequently used in ecology.

We first describe a superposition analogy showing its heuristic value for associative research, surrogate learning and inductive thinking [15, 16]. Following from this first potential, the second relates to the generation of new questions and hypotheses [17, 18]. This paper exploits both opportunities to show that the superposition analogy has potential to indicate recognized shortcomings of currently applied species concepts based on vegetative morphology. We provide a novel theoretical discussion about eco-evolutionary implications, particularly pertaining to cryptic species. Although the potential for reconciling classical and quantum logics for better understanding ecological systems may be enormous, our discussion about eco-evolutionary implications is purposefully preliminary, speculative and theoretical until suitable data for empirical testing become available.

Superposition analogy

In this paper superposition is used in the form of an analogy, a useful approach for relating quantum mechanics with classical systems [4, 5]. That is, our analyses are not based on a quantum framework per se but rather use the superposition analogy to show parallels between macrosystems and the quantum world. We borrow the concept from quantum physics to inform ecology and acknowledge that our results do not have any implication for quantum patterns and processes.

We emphasize that the analogous use of superposition does not allow for a mechanistic 1:1 translation of features from the subatomic world to macrosystems. In quantum mechanics, superposition arises from the

simultaneous occurrence of different states of an object.

These states comprise potentialities in an intangible reality and only one of these potentialities will eventually actualize and manifest upon measurement while all other potentialities become “annihilated”, according to the Copenhagen interpretation of quantum physics. In a simple example, an electron can simultaneously have a left and right spin before measurement, but once measured either the left or right spin will become apparent. In contrast, individuals in ecological communities are manifested, tangible entities that may be detected by sampling rather than inexistent, unperceived potentialities. Notwithstanding, situations may arise in ecology where patterns are phenomenologically reminiscent of superposition.

Consider clonal species (e.g., protists, bacteria) that can occur simultaneously within and across habitats. In this example superposition is intuitive: identical strains of an organism that reproduce by binary fission have simultaneous spatial distributions. This is a simple and straightforward example – akin to Occam's razor – of superposition informing species distributions. In this paper we focus on a less intuitive application of superposition. Our interest is in the analogous use of superposition to explain the counterintuitive pattern that even within a single ecosystem a single species can operate at more than one spatiotemporal scale.

Spatiotemporal scaling is a fundamental aspect of the complexity inherent in ecosystems and explicitly accounted for in ecological resilience theory [19]. In the context of this paper, accounting for scaling in statistical analyses is a critical first step towards detecting the illogical result of superposition that would go unnoticed if scale is not accounted for in analyses. Specifically, modeling results can reveal distinct ecological dimensions inherent in different spatiotemporal scales which differ substantially in their ecological structures and functioning [19]. Thus, detecting a modeling result that implies that a single species can operate simultaneously in independent, non-overlapping ecological niches would be at odds with ecological theory and common sense.

However, such a counterintuitive result may be informed by considering the superposition analogy. Combined with ecological resilience theory, the heuristic value of this analogy is to point out potential inconsistencies with the traditional species concept that can be demonstrated empirically through modeling. In the present case, we suggest that eco-evolutionary theories framed around cryptic species or syngens (ecologically differentiated and reproductively isolated lineages of taxa that conserve morphological similarity; [20, 21]) may inform such inconsistencies.

We envision that quantum superposition in ecological modeling, such as canonical ordination, including

redundancy analysis, is not at odds with classical scrutiny when interpretations discern genetically and ecologically differentiated subpopulations from morphological species (Fig. 1). Specifically, purposeful or deliberate lumping of populations of cryptic species into a community of a single morphological species shall manifest in a “smearing out” of such morpho-species across independent dimensions revealed by modeling. This leads to the paradoxical “both/and” scenario inspired by superposition (Fig. 1

creating orthogonal (statistically independent) canonical axes (or RDA axes) that are built from linear combinations of response variables that are simultaneously linear combinations of the explanatory variables. Ecologically these axes imply different structuring by environmental and biotic factors.

The RDA axes resolved by the models have often been correlated with the time series of individual taxa [31].

The result can be a classical “either/or scenario” in which a species correlates significantly with one axis in a significant model (Fig. 1). However, due to the nature of RDA, a situation may arise where a taxon can also be significantly correlated with more than one axis. Such a paradoxical “both/and” result may explain the neglect of such a scenario by ecologists. We specifically consider RDA for this study due to its adaptation to fit the premises of quantum theory. That is, we eliminate linear trends in time series [26] to create a quantum analogue of the classical RDA models. This approximation can be achieved through detrending the RDA models. This emulates quantum systems that deal with time less stringently than classical systems, meaning that the future can inform the past, which has been demonstrated in delayed-choice quantum erasure experiments [32]. Concomitantly, the RDA approach can mimic the particle aspect (e.g., electrons and photons) of the particle-wave duality in quantum systems, through analysis of species presence/absence patterns rather than biomass- or abundance-based data.

Materials and methods

We selected data of different taxonomic groups and ecosystems and used time series and spatial modeling based on redundancy analysis to showcase the superposition analogy in our modeling approach.

Data

For time series analyses we used two data sets. The first data set was obtained from the publicly available US Breeding Bird Survey (BBS) of North America, which contains avian community composition that is collected by qualified observers along georeferenced, permanent roadside routes across North America [33]. Along each approximately 39.5 km route, observers make 50 stops once every 0.8 km and conduct point-count surveys. During each survey, observers record for three minutes the abundance of all bird species that are acoustically or visually detected within a 0.4 km radius. Surveys start thirty minutes before local sunrise and last until the entire route is finished. To increase uniformity in probability of bird detection, surveys are conducted only on days with little or no rain, high visibility, and low wind.

For this study, we selected the South Central Plains as an example of a terrestrial ecosystem. We averaged three transects spanning the latitudes 31.8 to 33.4, which were

consistently sampled between 1968 and 2014 (47 years of data). We removed all aquatic species from the families Anseriformes, Gaviiformes, Gruiformes, Pelecaniformes, Phaethontiformes, Phoenicopteriformes, Podicipediformes, Procellariiformes, and Suliformes from analyses because of known negative observation biases for waterfowl compared with terrestrial avian families [34, 35]. We also removed hybrids and unknown species, and we condensed subspecies to their respective species following [36].

Our second data set for time series analysis contains phytoplankton community data from lake Stensjön, a small (surface area 0.57 km²), nutrient-poor, circumneutral lake located in the northern boreal forest biome of central Sweden (long: 14.77, lat: 56.45). Lake Stensjön is included in the Swedish National Lake Monitoring Program, which was established in the 1970s to assess the impact and recovery of anthropogenic acidification [37].

The monitoring program is overseen and regulated by the Swedish Agency for Marine and Water Management (HaV: <https://www.havochvatten.se/en>). Data are open access and available: <http://miljodata.slu.se/mvm/>.

For this study we used data spanning the period 1992 to 2018. Integrated samples of phytoplankton were collected from 5 sites in the upper stratification layer of the lake (epilimnion) in August with a plexiglass tube sampler (2 m long, inner diameter 10 cm), pooled and preserved in Lugol's solution. Phytoplankton counts were made using an inverted light microscope following the modified Utermöhl technique commonly used in Scandinavia [38]. Taxa were usually identified to the species-level taxonomic unit.

For the spatial analysis we used an exhaustive set of littoral invertebrate community data from 105 lakes sampled in 2017 that were distributed across Sweden ([39]; Fig. 1). The studied lakes all belong to the Swedish National Lake Monitoring program (see above), are medium sized (area=0.03–14 km², mean=1.5 km²) and are considered least disturbed in terms of no impact from point sources of pollution and land-use [36]. Sampling and analyses protocols for invertebrates are certified and quality controlled through the Swedish Board for Accreditation and Conformity Assessment (SWEDAC; <http://www.swedac.se/en/>) and followed Swedish standards (SS-EN 27,828).

Invertebrates were collected from each lake in one wind-exposed, vegetation-free littoral habitat during late autumn. In the most northern lakes, sampling was conducted between September and November to achieve similar seasonal conditions across surveys. Five replicate samples were taken, using standardized kick sampling with a hand net (0.5 mm mesh size). For each sample, the bottom substratum was disturbed for 20 s along a 1 m stretch of the littoral zone at a depth of ~0.5 m.

Invertebrate samples were preserved in 70% ethanol (estimated final concentration) in the field and processed in the laboratory by sorting against a white background with 10× magnification. Invertebrates were identified to the finest taxonomic unit possible and counted using dissecting and light microscopes.

Redundancy analysis

We carried out two time series analyses, one for the bird and the other for the phytoplankton data set, and one spatial analysis for the invertebrate data set based on RDA. All statistical analyses were carried out in R 3.6.1 [40] using packages *vegan* [41], *adespatial* [42], *ade4* [43] and *quickMEM* [44].

For the time series analyses, Moran Eigenvector Maps (MEM) [26, 27, 45], which comprise a set of orthogonal temporal variables, were obtained through the conversion, akin to a Fourier transformation, of the time vectors of the bird and phytoplankton time series. These time vectors consisted of 47 steps (sampling years) between years 1968 and 2014 for birds and 24 steps between 1992 and 2018 for phytoplankton, respectively. As a result of the Fourier transformation, these temporal

MEM variables take on the shape of sine waves of different wavelengths, which allows assessing fluctuation patterns at different inter-annual and interannual scales in the bird and phytoplankton data. These MEM variables are then used as explanatory variables to model temporal relationships in the bird and phytoplankton incidence data using redundancy analysis (RDA) [24].

Using forward selection, RDA selects significant MEM variables that best explain the temporal structures

prerequisite. Following [46], we used Spearman rank correlation analysis to relate incidence data of individual bird, phytoplankton and invertebrate taxa with the modeled patterns (lc scores) associated with the RDA axes of the respective models. This allowed us to assess how prevalent superposition is in the analyzed communities relative to correlations of species with only a single or no axis across species and data sets. Taxa that do not show significant correlations are considered to be stochastic because their dynamics are unrelated to the deterministic gradients revealed by RDA and are thus random with respect to these specific analyses [29, 36]. However, such species have traditionally been down-weighted by ecologists using RDA, although they may be relevant for understanding important ecological facets such as adaptive capacity or resilience [47]. We report the prevalence of all these fractions (species correlating with more than one axis [i.e. those showing superposition], those correlating with only one axis, and those not correlating with any axis) for broadest contextualization and comparisons of our results.

Results

RDA models

Time series analysis for birds and phytoplankton and spatial analysis for invertebrates revealed significant models

for all organism groups, although the proportion of variance of the minimum models (adjusted R^2) explained was low (birds: 0.08, phytoplankton: 0.2, invertebrates: 0.04). All models resolved more than one significant temporal or spatial dimension (RDA axes), thereby building the necessary basis for testing for the “both/and” superposition scenario. These models were manifested in 2 and 4 significant temporal dimensions for birds and phytoplankton, respectively; the spatial model for invertebrates revealed two significant spatial patterns (Fig. 3). The time series models for birds and phytoplankton showed broader scale (i.e. slower) patterns of community fluctuations associated with RDA 1 relative to the other RDA axes, which displayed faster community turnover (Fig. 3). Similar patterns were found in the spatial analysis for invertebrates, which displayed slightly more broad-scale patterns with RDA 1 relative to RDA 2.

Correlation analyses

Spearman rank correlation analyses revealed a low prevalence of superposition in the RDA models. For birds, 5 species (4.5% of the total number of species ($n=110$)) correlated with both significant axes of the RDA model, thus showing superposition in terms of simultaneously displaying independent temporal dynamics (Table 1). The remaining species correlated either with RDA 1 (9%) or

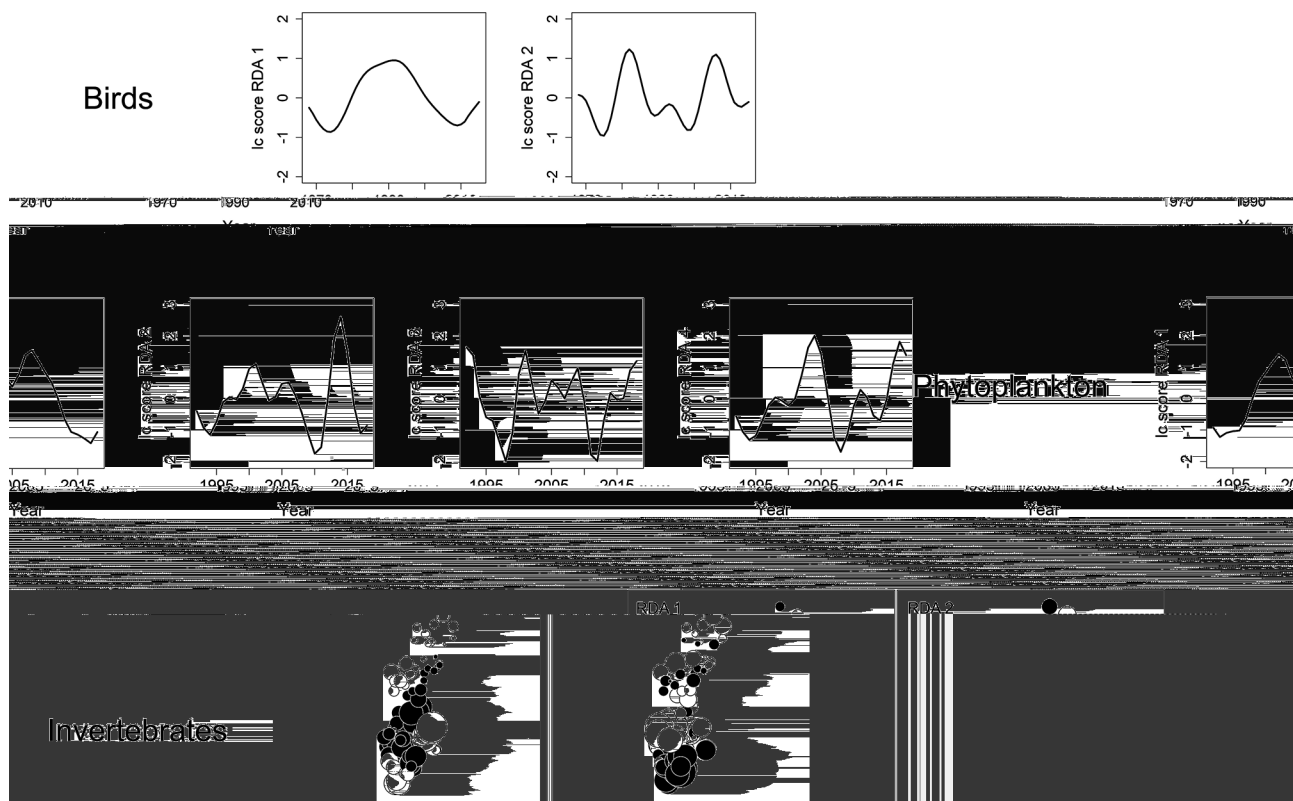


Fig. 3 Linear combination score plots associated with significant RDA axes in the MEM-RDA analyses showing statistically independent temporal patterns for birds and phytoplankton and spatial patterns for littoral invertebrates

Table 1 Results from Spearman rank correlation analysis testing for superposition. Shown are significant spearman rank correlation coefficients (ρ) at a significance level $P \leq 0.05$. Abbreviations: na, not applicable; --, not significant. Note that signs of correlations are considered not relevant for assessing superposition

	RDA 1	RDA 2	RDA 3	RDA 4
Birds				
<i>Leuconotopicus borealis</i>	-0.346	-0.346	na	na
<i>Protonotaria citrea</i>	0.323	-0.337	na	na
<i>Sayornis phoebe</i>	0.478	0.291	na	na
<i>Setophaga dominica</i>	0.367	0.387	na	na
<i>Troglodytes aedon</i>	-0.403	0.379	na	na
Phytoplankton				
Chroococcales (undefined)	0.492	0.435	--	--
<i>Dinobryon</i> sp.	0.385	0.521	--	--
<i>Dinobryon bavaricum</i> var. <i>vanhoe enii</i>	0.469	0.632	--	--
<i>Monoraphidium minutum</i>	0.445	0.484	--	--
<i>Cryptomonas marssonii</i>	0.542	--	0.494	--
<i>Staurodesmus triangularis</i> var. <i>limneticus</i>	0.455	--	0.428	--
<i>Chrysochromulina</i> sp.	0.453	--	--	-0.502
Undefined flagellates	-0.381	--	--	0.381
<i>Gymnodinium uberrimum</i>	0.499	--	--	-0.439
<i>Monoraphidium dybowskii</i>	0.415	--	--	0.495
<i>Parvodinium inconspicuum</i>	0.428	--	--	0.514
<i>Snowella atomus</i>	0.412	--	--	0.435
<i>Staurostrum longipes</i>	-0.499	--	--	0.439
<i>Ceratium hirundinella</i>	--	0.412	--	0.629
<i>Tabellaria flocculosa</i> var. <i>flocculosa</i>	--	0.402	--	0.632
<i>Urosolenia longiseta</i>	--	0.412	--	0.629
<i>Chroococcus minutus</i>	--	0.689	--	0.392
<i>Rhabdogloea smithii</i>	--	--	-0.383	-0.424
<i>Quadrigula closterioides</i>	-0.402	-0.388	--	-0.468
<i>Pseudokephyron</i> sp.	0.402	0.388	--	0.468
<i>Synura</i> sp.	0.391	0.445	--	0.401
Invertebrates				
<i>Cyrrus insolutus</i>	0.205	0.239	na	na
<i>Paratanytarsus</i> sp.	-0.356	-0.298	na	na
<i>Sialis lutaria</i>	0.217	0.194	na	na

RDA 2 (4.5%) or were not significantly correlated with any axis (stochastic species) (82%). For phytoplankton, 21 species (7% of the total number of species ($n=310$)) showed superposition (Table 1). The remaining species correlated either with RDA 1 (11%), RDA 2 (6%), RDA 3 (4%) or RDA 4 (3%), or were not significantly correlated with any axis (stochastic species: 68%). For invertebrates, 3 species (3% of the total number of species ($n=119$)) showed superposition in terms of occurring simultaneously in the two spatial dimensions (planes) resolved by the RDA (Table 1). The remaining species correlated either with RDA 1 (26%) or RDA 2 (6%) or were stochastic species (66%).

Discussion

The results of this study support our first goal showing that quantum superposition can occur in analogous form in RDA models across distinct organism groups and

ecosystems. For our research question it was relevant, first and foremost, to detect superposition independent of how many species fit the scenario. We therefore do not see the low prevalence of superposition as an inference limitation. We are aware that low incidence of superposition may be due to the nature of the approach and its adaptation to test for quantum superposition. RDA is robust against Type 1 statistical errors [48], which suggests that the modeling results do not confound patterns with random noise. However, RDA frequently yields limited explanatory power due to the nature of correlative analysis in which residual variation can be introduced due to the accumulation of noise resulting from sampling, survey designs, ecosystem history and system-intrinsic variation [49]. The low amount of variance explained can also be attributed to the correction of R^2 -values by the number of explanatory variables for obtaining appropriate models [27]. Furthermore, detrending models is an

additional source of loss of variance explained [26, 50]. Finally, using incidence rather than quantitative data of species (abundances, biovolumes) in the RDA models and correlation analyses might have further contributed to a decreased statistical performance of our analysis and thus the detection of low prevalence of superposition. As a result, the estimates of prevalence of superposition in our study might be conservative. However, the overall partitions of species correlating with a single or no axis (stochastic species) matches results from previous times series [31] and spatial analysis [46], which suggests that incidences of superposition in RDA models might be generally low. Notwithstanding, we acknowledge that different data sets and analysis designs could have probably resulted in different prevalence patterns of superposition but we purposefully traded off statistical performance in favor of adapting the approach to specifically account for premises of quantum mechanics (“free flow of time”, species as particles) in our analyses. Due to the properties of RDA, the temporal redundancy patterns of species incidences most likely lead to the showed superposition. However, we cannot ascertain why superposition was limited to the specific taxa in this study. In the absence of ecological variables mediating these patterns attributing taxon-specific responses is currently not possible.

Our study is based on the analogous use of quantum superposition in the RDA models. Using such an analogy clearly prevents a mechanistic 1:1 extrapolation, application and interpretation of superposition in ecology. In ecological communities, individuals are measured and observable entities. They do not comprise abstract potentialities that can occur in superposition in an intangible reality. Superposition, while being the norm for subatomic entities is therefore clearly at odds with how we generally understand and interpret macroscopic situations. The analogy in this study therefore, rather than comprising superposition in the sense of quantum physics, builds on a pattern akin to superposition that is generated by the RDA models. In this specific approach, this is due to the linear associations between the species that are redundant with and explained by a set of temporal or spatial predictor variables. The patterns of superposition of species as measured entities in the RDA models therefore results entirely from a mathematical and statistical procedure. That is, from creating orthogonal RDA axes that are built from linear combinations of species and explanatory variables [30].

Despite the discrepancies between the nature and manifestation of superposition in quantum mechanics and in the RDA models, the analogous use of superposition for the purpose of this study is useful for inspiring classical ecology theory, as has been shown for social [4] and economic [5] systems. However, we acknowledge that our study is cross-disciplinary in a sense that quantum

mechanics inspires ecology rather than the other way round. This brings us to the second goal of this study: reconciling the manifested quantum phenomena in the RDA models with eco-evolutionary patterns that are consistent with classical ecological logic. We will base our discussion on genetically differentiated subpopulations of taxa with near-identical morphology, i.e. cryptic species or syngens [20].

Unique vegetative morphology can independently emerge at different times during evolution, showing that morphology is not necessarily a marker of a monophyletic group or taxonomic species [51]. That is, species which share vegetative morphology (i.e., “morphological species”) can consist of reproductively isolated and ecologically differentiated subpopulations. Only members of a specific subpopulation (cryptic species or syngens) are compatible for mating, thereby fitting the biological species concept [52, 53]. Cryptic species are remarkably diverse among microscopic organisms [54, 55], but are also widespread in animals and plants [56–60]. With the continued development of molecular techniques even more cryptic species across organism groups, ecosystems and biomes are likely to be discovered. Despite the potential arising for biodiversity research, ecological and evolutionary factors that shape or are shaped by cryptic species have received limited research attention [21]. Given this dearth of information in the literature and the lack of data for empirical testing the following discussion about the reconciliation of our results with quantum theory is theoretical and aimed at stimulating future research.

Fišer et al. [21] consider cryptic species as a window for a paradigm shift of the species concept. Our results suggest that such a consideration is warranted. Ecological research is strongly biased towards morphological species routinely evaluated in traditional taxonomic studies, rather than using cryptic species complexes concealed in a morphological species identified by molecular methods and experiments that ascertain their ecological distinctness. RDA has strong potential to identify groups of taxa with similar ecological patterns resulting from intrinsic (e.g., nutrients, temperature, biological interaction) and extrinsic (e.g., habitat connectivity, spatial patterns) factors and temporal change that affect ecological communities. The multiscale nature of RDA also allows distinguishing between different deterministic patterns and can therefore indicate ecologically distinct groups of species in a community. RDA therefore has strong potential to test explicitly for the complex and non-linear factors that shape ecosystems, including cryptic species complexes, across different scales of space and time [61–64].

While a wealth of different study and analysis designs may have the potential to reveal the ecological distinctness of cryptic species, the usefulness of RDA per se as

a quantitative method for assessing such ecological differentiation of cryptic taxa needs further evaluation. However, our study suggests that RDA serves as a useful heuristic, inspired by quantum physics, to inform about the limitations when not accounting for ecological differentiation of subpopulations or when relevant environmental factors are not included in the analysis. Specifically, using traditional taxonomic surveys based on vegetative morphology may mask the differentiation of cryptic species in the analysis and add noise. The heuristic value of RDA therefore resides in indicating shortcomings of traditional analyses in the form of morphological taxa becoming “smeared out” across RDA axes, resulting in the counterintuitive and illogical superposition pattern. As quantum superposition becomes allegorical of the limitations of species concepts based on morphological criteria reported in the literature [21].

There is increasing evidence that cryptic species not only differ at the genomic level but also in environmental optima mediated by different functional traits [65]. This suggests ecological distinctness and the occupation of different ecological niches, which may manifest with the association of cryptic species with different axes in the RDA heuristic (Fig. 1B, right panel). We currently lack the exhaustive genomic functional trait data with sufficient spatial and temporal resolution for testing to what extent the eco-evolutionary analysis of cryptic species fits our heuristic. Future research using such extensive data sets, and likely simulation studies, may be useful for this purpose. The analogous use of quantum physics in such research may inspire hitherto unrealized potential for novel research in ecology.

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Author contributions

DGA conceived the study and wrote the paper. HBF-P analyzed the data, prepared the Figures and commented on an advanced paper draft. Both authors have approved submission of the paper.

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Data availability

Published data and R code can be found in the Zenodo archive at <https://doi.org/10.5281/zenodo.5880132>.

Declarations

Ethics approval and consent to participate

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Consent for publication

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Competing interests

The authors declare no competing interests.

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References

- Rae A. Quantum physics: illusion or reality. Cambridge Univ. Press; 2004.
- Kastner RE. Understanding our unseen reality – solving Quantum riddles. Imperial College Press; 2015.
- Wendt A. Quantum Mind and Social Science. Cambridge Univ. Press; 2015.
- Zohar D, Marshall I. Quantum Society. William Morrow Paperbacks; 1995.
- Orrell D. Quantum Economics: the Science of Money. Icon Books Ltd; 2018.
- Bull JW, Gordon A. Schrödinger's microbe: implications of coercing a living organism into a coherent quantum mechanical state. *Biol Philos*. 2015;30:845–56.
- Rodríguez RA, Herrera AM, Riera R, Santander J, Miranda JV, et al. Distribution of species diversity values: a link between classical and quantum mechanics in ecology. *Ecol Modell*. 2015a;313:162–80.
- Real R, Márcia Barbosa A, Bull JW. Species distributions, quantum theory, and the enhancement of biodiversity measures. *Syst Biol*. 2016;66:453–62.
- Rodríguez RA, Herrera AM, Riera R, Escudero CG, Delgado JD. Empirical clues about the fulfillment of quantum principles in ecology: potential meaning and theoretical challenges. *Ecol Modell*. 2015b;301:90–7.
- Bull JW. Quantum conservation biology: a new ecological tool. *Cons Lett*. 2015;8:227–9.
- Bagarello F. Quantum concepts in the Social, Ecological and Biological sciences. Cambridge Univ. Press; 2019.
- Alrøe HF, Noe E. Sustainability assessment and complementarity. *Ecol Soc*. 2016;21(1):30.
- Schrödinger E. Die gegenwärtige Situation in der Quantenmechanik. *Naturwissenschaften*. 1935;23(50):844–9.
- Romero-Isart O, Juan ML, Quidant R, Cirac JJ. Toward quantum superposition of living organisms. *New J Phys*. 2010;12(3):033015.
- Magnani L, Nersessian N, Thagard P. Model-based reasoning in Scientific Discovery. Dordrecht, The Netherlands: Kluwer; 1999.
- Swoyer C. Structural representation and surrogate reasoning. *Synthese*. 1991;87:449–508.
- Hesse M. The structure of scientific inference. London, UK: Macmillan; 1974.
- Holyoak K, Thagard P. Mental Leaps. Analogy in Creative Thought: Bradford, Cambridge, MA, USA; 1995.
- Angeler DG, Allen CR. Quantifying resilience. *J Appl Ecol*. 2016;53:617–24.
- Sonneborn TM. Breeding systems, reproductive methods, and species problems in Protozoa. *The Species Problem*. 1957;50:155–324.
- Fiser C, Robinson CT, Malard F. Cryptic species as a window into the paradigm shift of the species concept. *Mol Ecol*. 2018;27:613–35.
- Rao CR. The use and interpretation of principal component analysis in applied research. *Indian J Stats Series A*. 1964;26(4):329–58.
- van den Wollenberg AL. Redundancy analysis. An alternative for canonical correlation analysis. *Psychometrika*. 1977;42(2):207–19.
- Legendre P, Legendre LF. Numerical ecology. Elsevier; 2012.
- Borcard D, Gillet F, Legendre P. Canonical ordination. Numerical ecology with R. Cham: Springer; 2018. pp. 203–97.
- Baho DL, Futter MN, Johnson RK, Angeler DG. 2015. Assessing temporal scales and patterns in time series: Comparing methods based on redundancy analysis. *Ecol Complexity*. 2015;22:162–168.
- Peres-Neto PR, Legendre P, Dray S, Borcard D. Variation partitioning of species data matrices: estimation and comparison of fractions. *Ecology*. 2006;87:2614–25.
- Dray S, Legendre P, Peres-Neto PR. Spatial modeling: a comprehensive framework for principal coordinate analysis of neighbor matrices (PCNM). *Ecol Modell*. 2006;196:483–93.
- Dray S, Péliissier R, Couteron P, Fortin MJ, Legendre P, Peres-Neto PR, Bellier E, et al. Community ecology in the age of multivariate multiscale spatial analysis. *Ecol Monogr*. 2012;82(3):257–75.
- Buttigieg PL, Ramette A. A guide to statistical analysis in Microbial Ecology: a community-focused, living review of multivariate data analyses. *FEMS Microbiol Ecol*. 2014;90:543–50.
- Baho DL, Drakare S, Johnson RK, Allen CR, Angeler DG. Similar resilience attributes in lakes with different management practices. *PLoS ONE*. 2014;9(3):e91881.
- Kim YH, Yu R, Kulik SP, Shih Y, Scully MO. Delayed choice quantum eraser. *Phys Rev Lett*. 2000;84(1):1–5.

33. Sauer JR, et al. The North American breeding Bird Survey, results and analysis 1966–2015 (Version 2.07.2017). Laurel, MD: USGS Patuxent Wildlife Research Center; 2017. <https://www.pwrc.usgs.gov/bbs/>
34. Roberts CP, Allen CR, Angeler DG, Twidwell D. Shifting avian spatial regimes in a changing climate. *Nat Clim Change*. 2019a;9:562–6.
35. Roberts CP, Twidwell D, Angeler DG, Allen CR. How do ecological resilience metrics relate to community stability and collapse? *Ecol Ind*. 2019b;107:105552.
36. Angeler DG, Roberts CP, Twidwell D, Allen CR. The role of rare avian species for spatial resilience of shifting biomes in the Great Plains of North America. *Front Ecol Evol*. 2022;10:849944.
37. Fölster J, Johnson RK, Futter MN, Wilander A. The Swedish monitoring of surface waters: 50 years of adaptive monitoring. *Ambio*. 2014;43(1):3–18.
38. Olrik KP, Blomqvist P, Brettum P, Cronberg G, Eloranta P. Methods for quantitative assessment of phytoplankton in freshwaters, part I. Stockholm, Sweden: Swedish Environmental Protection Agency; 1989.
39. Fried-Petersen HB, Araya-Ajoy YG, Futter MN, Angeler DG. Drivers of long-term invertebrate community stability in changing Swedish lakes. *Glob Change Biol*. 2020;26(3):1259–70.
40. R Development Core Team. A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing; 2019.
41. Oksanen JF, Blanchet G, Friendly M, Kindt R, Legendre P, McGlinn D et al. vegan: Community Ecology Package. R package version 2.5-6. <https://CRAN.R-project.org/package=vegan>; 2019.
42. Dray S, Bauman D, Blanchet G, Borcard D, Clappe S, Guenard G, Jombart T et al. adespatial: Multivariate Multiscale Spatial Analysis. R package version 0.3-7. <https://CRAN.R-project.org/package=adespatial>; 2019.
43. Dray S, Dufour A. The ade4 Package: implementing the duality Diagram for ecologists. *J Stat Softw*. 2007;22(4):1–20.
44. Borcard D. Quick exploratory dbMEM analysis. Version 1.0.0. <https://raw.githubusercontent.com/zdealeveindy/anadat-r/master/scripts/NumEcolR2/quickMEM.R>; 2016.
45. Dray S. Moran's Eigenvector Maps and related methods for the spatial multi-scale analysis of ecological data. Available at: <https://cran.r-project.org/web/packages/adespatial/vignettes/tutorial.html>; 2019.
46. Angeler DG, Allen CR, Uden DR, Johnson RK. Spatial patterns and functional redundancies in a changing boreal lake landscape. *Ecosystems*. 2015;18(5):889–902.
47. Angeler DG, Fried-Petersen H, Allen CR, Garmestani A, Twidwell D, Chuang W, Donovan VM, Eason T, Roberts CP, Sundstrom SM, Wonkka CL. Adaptive capacity in ecosystems. *Adv Ecol Res*. 2019;60:1–24.
48. Angeler DG, Viedma O, Moreno J. Statistical performance and information content of time lag analysis and redundancy analysis in time series modeling. *Ecology*. 2009;90(11):3245–57.
49. Leibold MA, Ecomomo EP, Peres-Neto P. Metacommunity phylogenetics: separating the roles of environmental filters and historical biogeography. *Ecol Lett*. 2010;13:1290–9.
50. Borcard D, Legendre P, Avois-Jacquet C, Tuomisto H. Dissecting the spatial structure of ecological data at multiple scales. *Ecology*. 2004;85:826–1832.
51. Angeler DG, Schagerl M, Coleman AW. Phylogenetic relationships among isolates of *Eudorina* species (Volvocales, Chlorophyta) inferred from molecular and biochemical data. *J Phycol*. 1999;35:815–23.
52. Mayr E. The bearing of the new systematics on genetical problems. *Adv Genet*. 1948;2:205–37.
53. Manhart JR, McCourt RM. Molecular data and species concepts in the algae. *J Phycol*. 1992;28:730–7.
54. Fenchel T, Esteban GF, Finlay BJ. Local versus global diversity of microorganisms: cryptic diversity of ciliated protozoa. *Oikos*. 1997;80:220–5.
55. Schagerl M, Angeler DG, Coleman AW. Intraspecific phylogeny of *Pandorina morum* (Volvocales, Chlorophyta) inferred from molecular, biochemical and traditional data. *Eur J Phycol*. 1999;34:87–93.
56. Baker RJ, Bradley RD. Speciation in mammals and the genetic species concept. *J Mammalogy*. 2006;87(4):643–62.
57. Fernandez CC, Shevock JR, Glazer AN, Thompson JN. Cryptic species within the cosmopolitan desiccation-tolerant moss *Grimmia laevigata*. *Proc Natl Acad Sci*. 2006;103(3):637–42.
58. Bickford D, Lohman DJ, Sodhi NS, Ng PK, Meier R, Winker K, Ingram KK, Das I. Cryptic species as a window on diversity and conservation. *Trends Ecol Evol*. 2007;22(3):148–55.
59. Pfenninger M, Schwenk K. Cryptic animal species are homogeneously distributed among taxa and biogeographical regions. *BMC Evol Biol*. 2007;7:121.
60. Shneyer VS, Kotseruba VV. Cryptic species in plants and their detection by genetic differentiation between populations. *Russ J Genet Appl Res*. 2015;5:528–41.
61. Holling CS. Cross-scale morphology, geometry, and dynamics of ecosystems. *Ecol Monogr*. 1992;62:447–502.
62. Allen CR, Garmestani AS, Havlicek TD, Marquet PA, Peterson GD, Restrepo C, Stow CA, Weeks BE. Patterns in body mass distributions: sifting among alternative hypotheses. *Ecol Lett*. 2006;9(5):630–43.
63. Allen CR, Angeler DG, Garmestani AS, Gunderson LH, Holling CS. Panarchy: Theory and Applications. *Ecosystems*. 2014;17(4):578–89.
64. Nash KL, Allen CR, Angeler DG, Barichev C, Eason T, Garmestani AS, Graham NAJ, Granholm D, Knutson M, Nelson RJ, Nystrom M, Stow CA, Sundstrom SM. Discontinuities, cross-scale patterns and the organization of ecosystems. *Ecology*. 2014;95(3):654–67.
65. Baillet B. New methods for improving water management - Exploring the role of diatoms in ecosystems. Doctoral Thesis No. 2021:55, Faculty of Natural Resources and Agricultural Sciences, Swedish University of Agricultural Sciences; 2021. Available at: https://pub.epsilon.slu.se/25042/1/baillet_b_210826.pdf.

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